

## Pattern of functional diversity along the elevation gradient in the dry evergreen Afromontane forest of Hararghe Highland, Southeast Ethiopia

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Understanding plant species distribution patterns along environmental gradients is fundamental to managing ecosystems, particularly when habitats are fragmented due to intensive human land use pressure. The variation pattern of functional diversity of plant communities along the elevation gradient in the Dindin dry evergreen Afromontane forest was tested. Fifty four plots of 20 x 20 m (400 m<sup>2</sup>) were established at 200 m intervals starting 2,300–2,900 m a. s. l. and woody species composition, and environmental variables were recorded. Nine functional diversity indices based on functional distances were employed to estimate functional diversity. The mixed effect model was used to determine the effect of elevation, aspect and slope on functional diversity indices. The results showed that functional diversity in communities varied greatly; functional diversity revealed a decrease with increasing elevation and a “humped” pattern, with peak diversity appearing at middle elevation. Functional diversity was significantly correlated with elevation, slope, and aspect. Functional diversity was significantly correlated with species richness and evenness. Environmental filtering was important to the functional diversity pattern; the nine indices were all successful in the analysis of functional diversity in the plant community with different effectiveness, and modified functional attribute diversity, plot-based functional diversity, community based functional diversity, functional richness, and community weight mean of woody density performed better than the other four indices in this study.

**Keywords:** environmental filters; community weight mean; mixed effect model; wood density; specific leaf area.

### Introduction

Afromontane forests are distributed across the African highlands (Linder, 2014; Abiem et al., 2020). Typically found above 1500 m in elevation, they extend from the Arabian Peninsula south along the East African Rift to the Drakensberg Mountains in the east. Despite this distribution, Afromontane forests have a unique fauna and flora and harbour a high proportion of endemic species (Burgess et al., 2007; Abiem et al., 2020). They have significant carbon stocks (Spracklen & Righelato, 2014), provide watershed protection (Schröter et al., 2005) and other important ecosystem services (Conti & Díaz, 2013). However, Afromontane forests tend to occur in areas of high human population pressure and are under intense threat from agricultural expansion, fire, and grazing (Cordeiro et al., 2007; Kebede et al., 2013). The Ethiopian highlands contribute more than half of the total land area covered by Afromontane forest, of which dry evergreen Afromontane forests occupy a large part (Friis et al., 2010). The topographic and elevation variation in Ethiopian landscapes affects the existence of different vegetation types and species diversity. A large topographic variation within small geographical areas makes the elevation gradient a unique position to study species diversity (Wana, 2009; Kebede et al., 2013). Long term monitoring of these forests is crucial for understanding how they function and for predicting their dynamics along environmental changes, vital information for sustainable management and preservation of the local and global services they provide.

Functional diversity is an important ecological attribute of a particular ecosystem (Mason et al., 2005; Conti & Diaz, 2013). In recent decades, functional diversity has received much attention from conservationists and ecologists (Zhang et al., 2017; Thakur & Chawla, 2019). Functional diversity is an important predictor of species in response to environmental changes and is used to describe specific functional adaptations of plants to different environmental factors (Mouchet et al., 2010; Mason & de Bello, 2013). It is now well established that many environmental factors, inclu-

ding topography (i.e. elevation, slope, and aspect), edaphic, and disturbance factors have been found to affect functional traits of plant communities (Pavoine & Bonsall, 2011; Costa et al., 2017). Notably, elevation is one of the most frequently stated problems with the functional trait variability of plant communities. In highland ecosystems, increasing elevation is usually related to lower temperatures and higher humidity (Solefack et al., 2018). Change in this climatic variable is a stress factor for plants, which affects plant diversity by selecting biological forms capable of existing at different elevation ranges (Klimes, 2003). In addition to elevation, slope and aspect also has a significant effect on functional diversity and is important to species composition, community structure, and soil development in ecosystems (Pavoine & Bonsall, 2011). Changes in slope aspects lead to changes in hours of sunshine, humidity, and temperature, which affect plant communities (Austrheim, 2002; Zhang et al., 2018).

Elevation found to be influencing functional diversity has been explored in several studies (Pavoine & Bonsall, 2011; Zhang et al., 2018). More recently, literature has emerged that offers contradictory findings about functional diversity pattern along an elevation gradient (Swenson et al., 2011; Zhang et al., 2014). For instance, in the Luquillo experimental forest, Swenson et al. (2011) reported that functional diversity increased along the elevation gradient. Kluge & Kessler (2011) suggested that environmental filtering as an important to functional diversity under stressful environmental conditions at low and high elevations, whereas interspecific competition was important to functional diversity at middle elevation. By contrast, Thakur & Chawla (2019) observed that functional diversity decreases with increasing elevation in the high altitude vegetation of the Western Himalaya. Likewise, de Bello et al. (2013) and Gazol et al. (2017) reported a decrease in functional diversity with increasing elevation in the Alps and South Urals, respectively. Zhang et al. (2014) reported a “humped” pattern showing peak functional diversity near the middle of the gradient. However, their study suggested a possible reason for the patterns of trait variation and functional diversity variation along the eleva-

tion gradient related to vegetation type and ecoregions. So far, little study has been done on functional diversity in relation to environmental gradient in sub-Saharan countries (Wana, 2009; Solefack et al., 2018). Most of the work on this subject has focused on the phylogenetic diversity of plant communities (Zhou et al., 2018), vegetation structure (Rutten et al., 2015), species richness pattern along elevation gradient (Tesfaye et al., 2010; Aynekulu et al., 2012; Kebede et al., 2013). This paper examines the relationship between functional diversity and environmental factors (i.e. elevation, slope and aspect) in a dry evergreen Afromontane forest. The response of functional diversity variation patterns to environmental change and the underlying causal mechanisms behind the patterns is not understood. To assist the future management and conservation of biological diversity, detecting the variation patterns of functional diversity and understanding the underlying causal mechanisms behind the patterns is essential. Therefore, the objective of this study is to describe and examine variation patterns of functional diversity of plant communities along an environmental gradient, to test the hypothesis that functional diversity decreases with increasing of elevation gradients due to environmental filtering, and to identify the key environmental variables affecting functional diversity in the studied forest.

### Material and methods

The study was conducted in Dindin dry evergreen Afromontane forest in the Hararghe Highlands, Southeast Ethiopia. The geographical lo-

cation of the study site lies between 40°10'40" to 40°18'50" E and 8°33'0" to 8°40'40" N with elevation ranges between 2,124 and 3,069 m a. s. l. and situated around 336 km southeast of Addis Ababa (Fig. 1). Due to the lack of long term climatic data for the study site, the climate estimator software tool, New LocClim (FAO, 2005; Grieser et al., 2006) was used to produce long term monthly rainfall and temperature data for the study site. The mean annual temperature (MAT) and mean annual precipitation (MAP) at the study site was estimated to be 25.6 °C and 804 mm/year, respectively. Precipitation at the study site has a bimodal distribution pattern with a long rainy season lasting from June to October and a short rainy season from April to May (Fig. 2). The soils of the study area developed from a wide range of parent materials, including volcanic and mixed limestone and sandstone over a Precambrian basement. Leptosols are most abundant soil types which are classified as Haplic and Lithic leptosols. Lithic leptosol covers are most extensive on high mountainous relief hills and parallel ridges and river gorges having very steep slopes (30–60%) (Tefera et al., 1996; Elias, 2016). According to Friis et al. (2010), the vegetation type in the study area can be categorized as Tropical dry evergreen Afromontane forest, which is characterized by a dry climate. Woody species such as *Afrocarpus falcatus*, *Maesa lanceolata*, *Allophylus abyssinicus* and *Vernonia myriantha* are dominant. Along the altitudinal gradient between 2,300 and 2,900 m a. s. l., a stratified systematic sampling of 54 plots separated by 200 m raises in elevation (2300, 2500, 2700 and 2900 m a. s. l.) were established at 200 m intervals along the four horizontal transects, the initial plot was selected randomly.

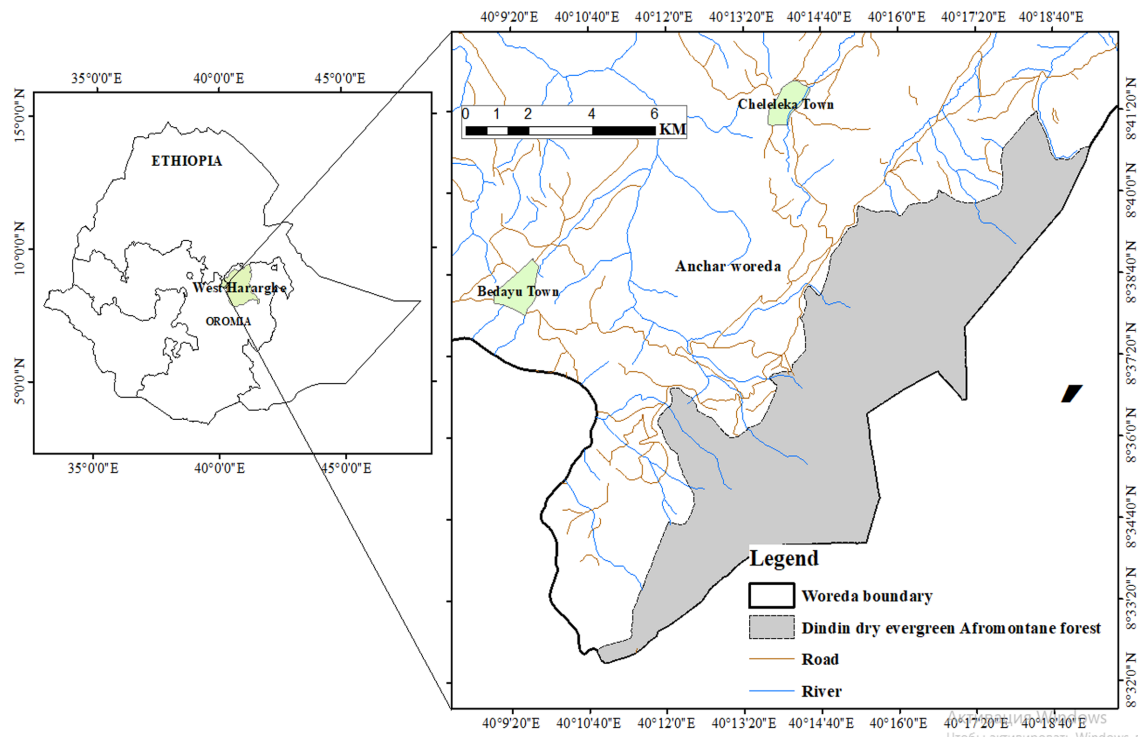


Fig. 1. Study area map

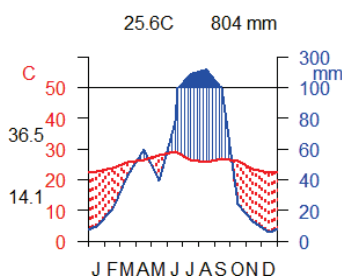


Fig. 2. Climate Diagram of study area (values generated using New\_LocClim. See text)

This approach was adopted for two main reasons: (i) there is a minimum sampling bias as all woody species in the study forest are represen-

ted; (ii) the effect of area (the decrease in area from bottom to top of a mountain) is minimized, as all forest areas are included. Moreover, this procedure gives a precise indication of species represented and is used to achieve better documentation of changes in species distribution patterns along an elevation gradient (Aynekulu et al., 2012; Arellano et al., 2016). The plot size of 20 x 20 m was used to record trees and woody lianas, in which three 5 x 5 m sub-plots were used to record shrubs. In each plot, height, diameter and individual abundance of trees and shrubs were recorded. Furthermore, all woody lianas with a dbh  $\geq 1$  cm at 1.30 m from the rooting point were also recorded, because this could provide a detailed assessment of its diversity (Arellano et al., 2016). The habits or growth forms of each species were also recorded. For the purpose of this study, trees are one stemmed woody plants with heights  $\geq 5$  m and shrubs are represented a multi stemmed woody plants with heights  $\leq 5$  m. Each species was identified in the field whenever possible, but in doubtful cases

vouchers were collected and pressed for further identification and confirmation at the Addis Ababa University national herbarium. Environmental variables such as elevation, slope, aspect and geographical location of each plot were recorded. The elevation was measured by using a GPS Garmin; whereas, the slope and aspect were measured by using a compass meter.

Two functional traits were selected to demonstrate plant species function in the community along an elevation gradient. Data on specific wood density was obtained from the Global Wood Density Database (Chave et al., 2009; Zanne et al., 2009). Wood density ( $g/m^3$ ) is the dry mass per unit volume of fresh wood, which is a fundamental characteristic of tree trunks (Muller-Landau, 2004) and it plays a major role in light acquisition strategies. Specific leaf area (SLA) was extracted from the TRY plant database (Kattge et al., 2011; Maire et al., 2015). SLA defined as leaf area ( $cm^2$ ) per unit of leaf dry mass (g). Average genus values for wood density and SLA was used when a particular species is missing.

There are various measurement methods for functional diversity based on species traits available in ecology. We used the following indices.

**Modified functional attributes diversity (MFAD):** MFAD measures the dispersion of species in the functional trait space (Schmera et al., 2009).

$$MFAD = \frac{\sum_{h=1}^N \sum_{k=1}^N d_{hk}}{N} \quad (1)$$

where  $d_{hk}$  is the distances between functional units  $h$  and  $k$  and  $N$  is the number of functional units.

**Functional diversity based on dendrogram (FDp and FDc):** Functional diversity based on dendrogram length is the sum of branch length of the functional dendrogram that can be calculated from a matrix of species functional traits (Petchey & Gaston, 2006). Gower's distance and group averaging clustering were used to calculate the dendrogram (Podani & Schmera, 2006). Gower distance is important in detecting changes in functional diversity along environmental gradients, as greater variations in trait values reflect increased niche differentiation (de Bello et al., 2012). FDp is a plot-based index which recalculates the dendrogram for each plot, FDc is a community based index (Casanoves et al., 2011).

**Functional divergence (FDiv):** FDiv refers to how abundance is distributed within the volume of the functional trait space (Villeger et al., 2008).

$$FDiv = \frac{\sum_{i=1}^S w_i (dG_i - \overline{dG}) + \overline{dG}}{\sum_{i=1}^S w_i |dG_i - \overline{dG}| + \overline{dG}} \quad (2)$$

where  $dG_i$  is the functional distance from species  $i$  to the gravity center of species that form the vertices of the convex hull, and  $\overline{dG}$  is the mean distance of the  $S$  species to the gravity center.  $w_i$  is the relative abundance of species  $i$ .

**Functional dispersion (FDis):** FDis is the weighted average distance in multidimensional trait space of individual species to the weighted centroid of all species, where the weights correspond to the relative abundances of the species (Laliberté & Legendre, 2010):

$$FDis = \sum_{i=1}^S w_i z_i \quad (3)$$

where  $w_i$  is the abundance of species ' $i$ ' and  $z_i$  is the distance of species ' $i$ ' to the weighted centroid  $c$ .

**Functional richness (FRic):** It refers to the amount of trait space filled with species in the community (Mason et al., 2005). FRic was computed as given below:

$$FRic = \frac{Fci}{Rc} \quad (4)$$

where  $Fric$  – the functional richness of functional character in community  $i$ ,  $Fci$  – the niche space filled by the species within the community,  $Rc$  – the absolute range of the character.

**Functional evenness (FEve):** Functional evenness measures the regularity of spacing between species in the trait space volume and also the evenness in the distribution of the species abundance. FEve calculated as (Pla et al., 2012):

$$FEve = \frac{\sum_{b=1}^{S-1} \min\left(PWE_b, \frac{1}{S-1}\right) - \frac{1}{S-1}}{1 - \frac{1}{S-1}} \quad (5)$$

where PEWb is the partial weighted evenness,  $S$  is species richness.

**Community weight mean (CWM) of traits:** Plot CWM of functional traits were calculated as the mean trait value of each genus weighted by the relative abundance of the genus in a given plot (Cavanaugh et al., 2014).

$$CWM(\text{trait}_x) = \sum p_i x_i \quad (6)$$

where CWM ( $\text{trait}_x$ ) is the CWM for  $X$  trait,  $p_i$  is the relative cover of species  $i$  in the community and  $x_i$  is the trait value for the species  $i$ .

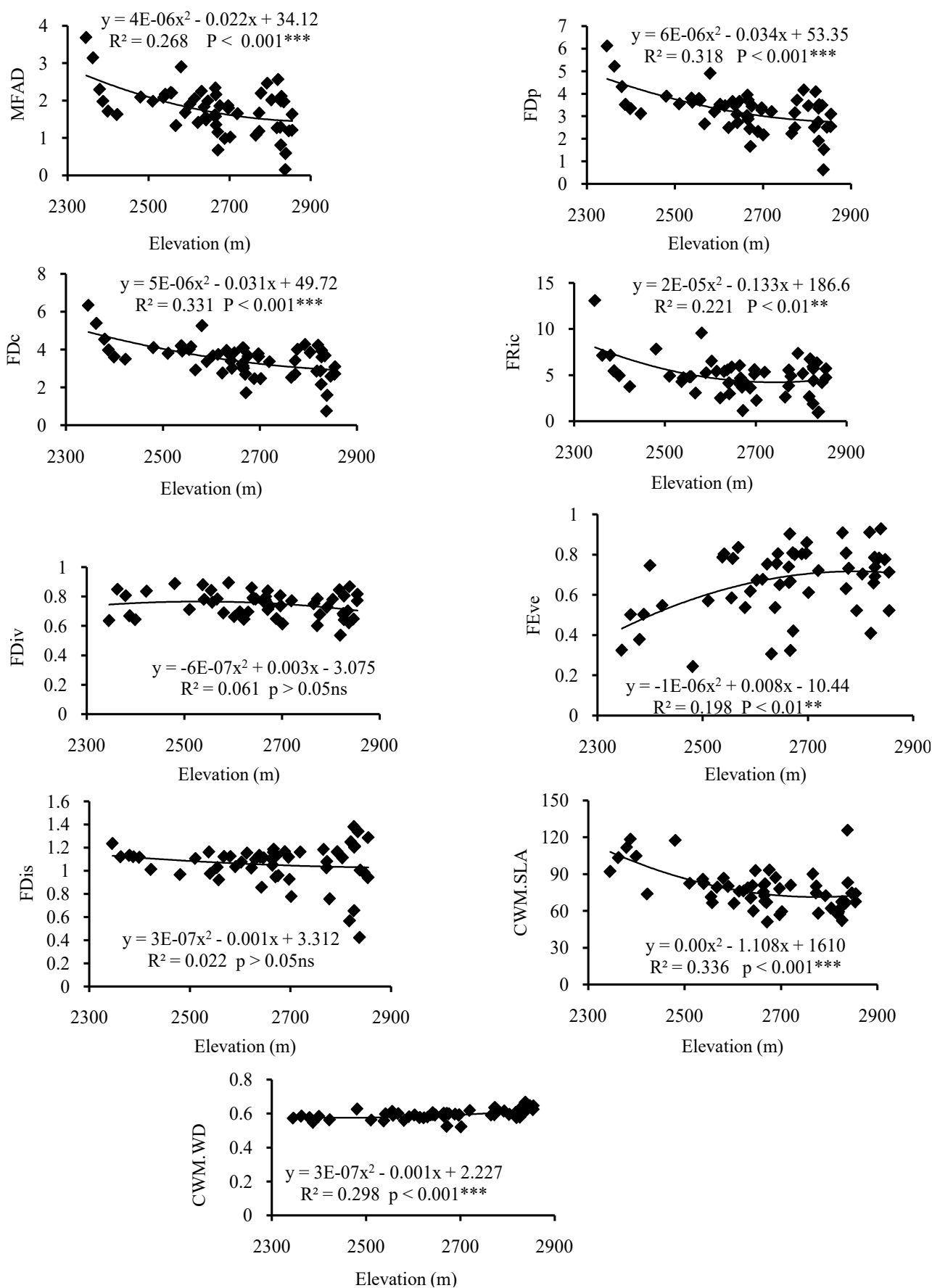
**Taxonomic diversity:** Three taxonomic diversity indices, species richness ( $S$ ), species heterogeneity (Shannon-Wiener index  $H'$ ), and one for species evenness (Pielou index  $E$ ), were used to calculate species diversity values (Magurran, 2004).

Spearman correlations were used to analyze relationships of functional diversity, environmental variables, and species diversity. Polynomial regressions were used to test the variation patterns of functional diversity along the elevation gradient. The functional diversity indices were computed by the FDiversity software (Casanoves et al., 2011). Multi-trait functional diversity indices were computed after standardization of trait values, while a single functional trait index was calculated without standardization. Standardization was performed because studied traits differ in several orders of magnitude and scale of measurement (Pla et al., 2012). Furthermore, to determine the effect of elevation, aspect and slope (fixed effects) on different functional diversity measures, a mixed effects model (Bates et al., 2015) with 'Type III analysis of variance with Satterthwaite's method' was used. To run the mixed effects model package 'lme4' (Bates et al., 2015) was used and variables were selected by 'backward selection' using 'cAIC4' package (Saeften et al., 2018). The significant effect of fixed factors was determined using "lmerTest" package (Kuznetsova et al., 2017) in R. All the functional diversity indices (response variables) were log-transformed prior to using in the mixed effects models to achieve normality assumption.

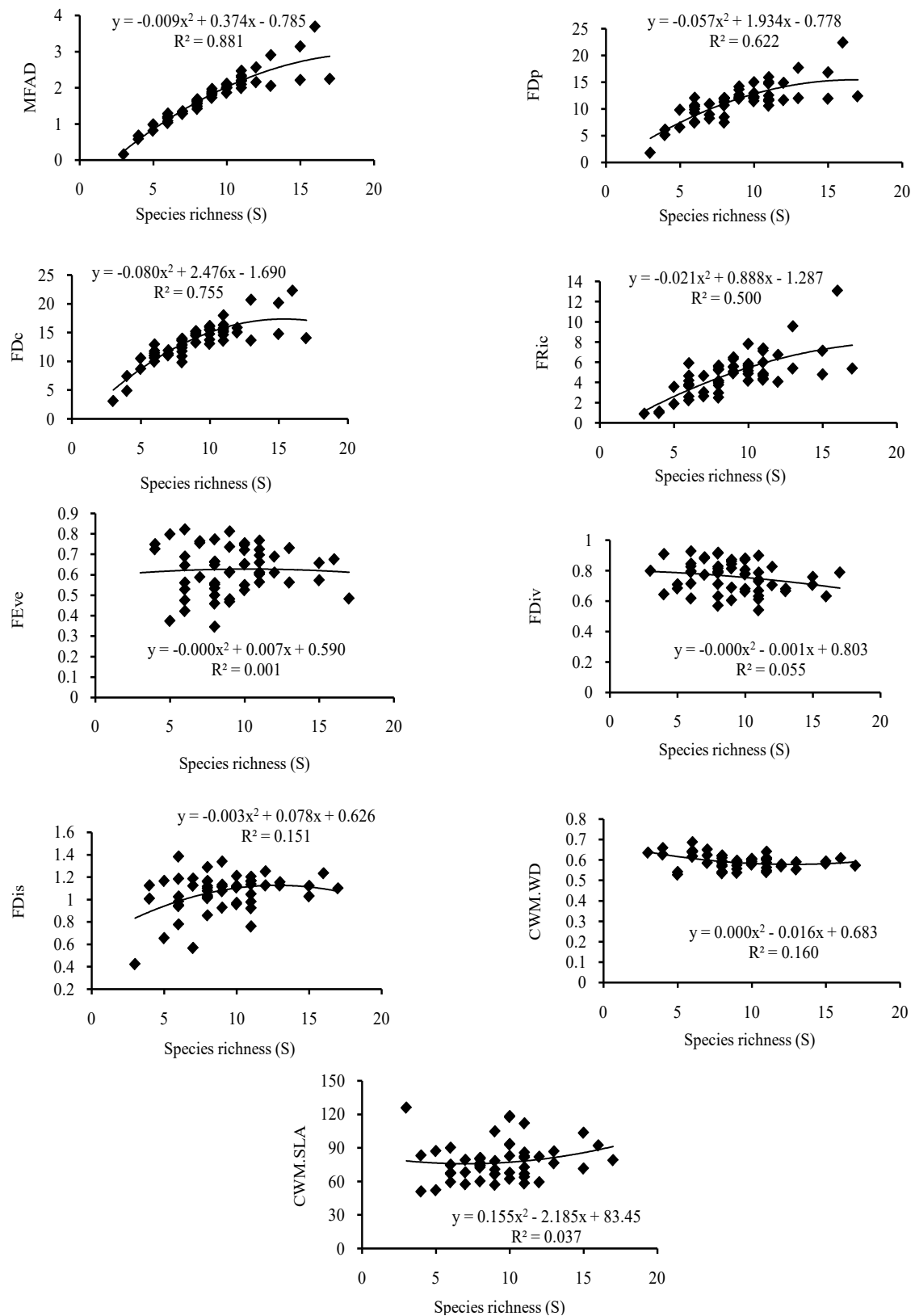
## Results

The functional diversity computed by the nine indices showed significant changes in the communities of the Dindin dry evergreen Afromontane forest. Modified functional attribute diversity (MFAD) varied between 0.152 and 3.683, Functional diversity plot based dendrogram index (FDp) between 0.609 and 6.111, Functional diversity community based dendrogram index (FDc) between 0.738 and 6.553, Functional richness (FRic) between 0.898 and 13.074, Functional evenness (FEve) between 0.249 and 0.804, Functional divergence (FDiv) between 0.535 and 0.892, Functional dispersion (FDis) between 0.161 and 1.233, and 1.936, community weight mean of wood density (CWM.WD) between 0.522 and 0.665 and community weight mean of specific leaf area (CWM.SLA) between 34.789 and 123.875. The effect of elevation gradient on most of the functional diversity indices was significant (Table 1). The functional diversity indices of woody species showed different elevation patterns, including decreases, increases, and increases then decreases. Out of the nine functional diversity indices, MFAD, FDp, FDc, FRic, FDis, and CWM.SLA provided a significant decreasing pattern (Fig. 3), whereas CWM.WD showed a slight increasing pattern with increasing elevation. The relationships between functional diversity and elevation were nonlinear (Fig. 3). In addition to elevation, other environmental variables such as slope and aspect were also a significant effect on spatial patterns of functional diversity (Table 1). There was a significant effect of elevation, slope and aspect on FDp and CWM.WD. The effect of elevation and slope was significant on MFAD and FDc, whereas slope and aspect were significantly affecting FDiv. The effect of elevation on FRic, FEve and CWM\_SLA was significant. In the case of FDis, CWM.SLA and FEve significant effects of aspect were found (Table 1). However, the effect of elevation on FDiv, FDis and FEve was not significant.

A functional diversity and CWM trait was significantly correlated with species richness, Shannon Weiner diversity, and evenness in communities of the studied forest. Seven out of nine functional diversity indices were positively correlated with species richness; two out of nine negatively correlated to species richness (Table 2). Six out of nine functional diversity indices revealed similar patterns, because they are correlated with species richness (Table 2 and Fig. 4). The relationships between functional diversity indices and species richness were non linear (Fig. 4).



**Fig. 3.** Change in functional diversity along an elevation gradient in the Dindin dry evergreen Afromontane forest: MFAD – modified functional attribute diversity, FDP – plot-based functional diversity, FDC – community based functional diversity, FDiv – functional divergence, FRic – functional richness, FEve – functional evenness, FDis – functional dispersion, CWM.WD – community weight mean of woody density ( $g/m^3$ ), CWM.SLA – community weight mean of specific leaf area ( $cm^2/g$ ); \*\*\*, \*\*, \* – significant effects, ns – not significant effects



**Fig. 4.** Relationships between functional diversity and species richness: MFAD – modified functional attribute diversity, FDp – plot-based functional diversity, FDc – community based functional diversity, FDiv – functional divergence, FRic – functional richness, FEve – functional evenness, FDIs – functional dispersion, CWM.WD – community weight mean of woody density ( $\text{g}/\text{m}^3$ ), CWM.SLA – community weight mean of specific leaf area ( $\text{cm}^2/\text{g}$ )

## Discussion

The results revealed that functional diversity varied significantly in plant communities of the Dindin dry evergreen Afromontane forest, which indicated that functional diversity was a suitable indicator for quan-

tifying relationships of function, and the environment in communities (Mason et al., 2005). In this study, the nine functional diversity indices used were all successful in demonstrating the quantity and variation pattern of functional diversity in communities, but their effectiveness was different (Mason et al., 2005; Zhang et al., 2014). In line with the hypothe-

sis, there were a significant decrease in functional diversity indices, particularly MFAD, FDP, FRic, and FDC showed a clear decrease with increasing elevation; which is attributed to increasing environmental filtering at high elevations (Kluge & Kessler, 2011; Gazol et al., 2017; Thakur & Chawla, 2019). These patterns suggest that functionally more similar species should coexist (i.e. convergence in plant functional trait) at higher elevations. This could be due to increased climatic severity, less competition for resources and increased positive interactions between species at high elevations.

**Table 1**

Summary of statistics from mixed effects model representing effect of elevation, aspect and slope gradient on functional diversity components with 'type III' analysis of variance and Satterthwaite's method

Functional diversity component	Effect	Estimate	d.f.	SE	P value
MFAD	elevation	-0.0011	43.846	0.0004	< 0.01**
	slope	-0.0064	40.740	0.0027	< 0.05*
FDP	elevation	-0.0009	43.673	0.0003	< 0.01**
	slope	-0.0039	41.345	0.0019	< 0.05*
	aspect	-0.4039	22.864	0.2281	> 0.05 <sup>ns</sup>
FDC	elevation	-0.0009	43.640	0.0003	< 0.001***
	slope	-0.0040	42.017	0.0018	< 0.05*
FRic	elevation	-0.0058	47.885	8.1620	< 0.01**
FDis	aspect	-0.3584	18.799	0.1543	< 0.05*
FDiv	slope	-0.0015	23.980	0.0008	> 0.05 <sup>ns</sup>
	aspect	-0.2558	43.980	0.1466	> 0.05 <sup>ns</sup>
FEve	aspect	-0.1824	43.381	0.1078	> 0.05 <sup>ns</sup>
CWM.WD	elevation	0.0001	43.120	0.0000	< 0.001***
	slope	-0.0005	42.450	0.0003	> 0.05 <sup>ns</sup>
	aspect	-0.1415	43.080	0.0646	< 0.05*
CWM.SLA	elevation	-0.0009	25.068	0.0002	< 0.001***
	aspect	0.2680	43.886	0.1402	> 0.05 <sup>ns</sup>

Notes: d.f. – degrees of freedom, SE – standard error; \*\*\*, \*\*, \* – significant effects, <sup>ns</sup> – not significant effects; MFAD – modified functional attribute diversity, FDP – plot-based functional diversity, FDC – community based functional diversity, FDiv – functional divergence, FRic – functional richness, FEve – functional evenness, FDis – functional dispersion, CWM.WD – community weight mean of woody density (g/m<sup>3</sup>), CWM.SLA – community weight mean of specific leaf area (cm<sup>2</sup>/g)

**Table 2**

Spearman correlation coefficients between the nine functional diversity indices and species richness, diversity and evenness in plant communities

Functional diversity component	Species richness (S)	Shannon Weiner diversity (H')	Evenness index (E)
MFAD	0.97***	0.87***	0.35**
FDP	0.94***	0.84***	0.34*
FDC	0.94***	0.85***	0.34*
FRic	0.66***	0.69***	0.41**
FEve	0.06 <sup>ns</sup>	0.14 <sup>ns</sup>	0.30**
FDiv	-0.28*	-0.26 <sup>ns</sup>	-0.10 <sup>ns</sup>
FDis	0.24 <sup>ns</sup>	0.42**	0.50***
CWM.WD	-0.36**	-0.37*	-0.18 <sup>ns</sup>
CWM.SLA	0.20 <sup>ns</sup>	0.34*	0.34*

Notes: \*\*\* – P < 0.001, \*\* – P < 0.01, \* – P < 0.05, <sup>ns</sup> – not significant; MFAD – modified functional attribute diversity, FDP – plot-based functional diversity, FDC – community based functional diversity, FDiv – functional divergence, FRic – functional richness, FEve – functional evenness, FDis – functional dispersion, CWM.WD – community weight mean of woody density (g/m<sup>3</sup>), CWM.SLA – community weight mean of specific leaf area (cm<sup>2</sup>/g).

Spasojevic et al. (2014) confirmed that species at high elevations are functionally more similar to each other and have lower functional diversity. In agreement with this study, Thakur & Chawla (2019) observed that FRic decreased with increasing elevation in the high altitude vegetation of the western Himalayas. Similarly, de Bello et al. (2013) and Gazol et al. (2017) reported a decrease in FRic with increasing elevation in the Alps and South Urals. In this study, functional dispersion (FDis) and functional divergence (FDiv) of woody species showed no significant decrease with increasing elevation; the underlying processes driving the patterns might be explained by the gradients of climate and habitat variables as well as biotic interactions such as competition (Chun & Lee, 2017). With high elevation vegetation, the decreasing pattern with increasing elevation of

FRic indicates less resource use at higher elevations, whereas, those of FDiv, and FDis indicate less niche differentiation, and niche complementarity at higher elevations (Kraft et al., 2008; Thakur & Chawla, 2019). Functional evenness showed a significant humped pattern. Similarly, Zhang et al. (2014) reported that functional evenness initially increases with increasing elevation and then decreases.

CWM.SLA showed a significant decrease with increasing elevation and may be attributed to the filtering effect of decreased temperature. Similar trends in CWM.SLA have been previously reported for Mount Hamaliya (Thakur & Chawla, 2019) and Mount Kilimanjaro (Costa et al., 2017). However, CWM.WD revealed a significant increase with increasing elevation. The dominant functional traits (high wood density and low specific leaf area) in the stressful areas are indicative of a stress tolerant life history strategy (Swenson & Enquist, 2007). The variation in patterns of traits and functional diversity along the elevation gradient might be attributed to vegetation type and ecoregions (Thakur & Chawla, 2019).

Other environmental variables slope and aspect also had significant effects on functional diversity. Slope and aspect are independent of elevation and are important to species composition, community structure, and soil development in ecosystems (Pavoine & Bonsall 2011). Zhang et al. (2018) noted that slope gradient is an essential factor affecting plant communities in alpine meadows, due to the physical and chemical characteristics of the soil and water content. Elliott & Kipfmüller (2010) suggest that seedling recruitment is strongly influenced by slope, aspect and is a potential complicating factor of future ecological dynamics on south facing slopes, particularly within the context of climate change.

There was significant correlation between functional diversity indices and species richness, Shannon, and Evenness in the studied forest. This suggests that functional diversity and species richness are interrelated with each other. A positive relationship of species richness and functional diversity indicates that each species has a relatively unique set of traits (Bu et al., 2014) and functional complementarity usually appears in some respects among different species. The positive relationship between species diversity and functional diversity in this study is consistent with previous studies (Caddotte et al., 2011; Zhang et al., 2014, 2018). These results build on existing evidence of the relationship between species diversity and functional diversity influenced by the sensitivity of the functional diversity metric to functional variation between species. Furthermore, the non linear relationships between functional diversity and species richness confirmed that they cannot be replaced with each other (Zhang et al., 2014). There are some changes in functional traits which may go beyond the change that can be explained only through taxonomic diversity variation (Ricotta & Moretti, 2008). The variation in species abundance and distribution can mainly explained through species diversity, however functional diversity can explain variations of species morphology, physiology, reproduction, ecology and phenology (Papanikolaou et al., 2011).

## Conclusions

Most of the functional diversity indices showed a decreasing pattern with increasing elevation in Dindin dry evergreen Afromontane forest, which indicates that elevation is the most important environmental filter which influences the species distribution, community composition and structure in the communities. Further, slope and aspect have a significant effect on the plant distribution and functioning of high elevation plant communities. Knowledge of the role of environmental filtering in changing community composition and functional diversity in the dry evergreen Afromontane forest is very important for understanding of diversity changes and ecosystem processes. Further studies should be conducted to monitor and predict the effects of climate change on the fragile dry evergreen Afromontane ecosystems which are mediated through soil moisture content.

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